

# From Elemental Sulfur to Hydrogen Sulfide in Agricultural Soils and Plants

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## Abstract

Sulfur is an essential element in determining the productivity and quality of agricultural products. It is also an element associated with tolerance to biotic and abiotic stress in plants. In agricultural practice, sulfur has broad use in the form of sulfate fertilizers and, to a lesser extent, as sulfite biostimulants. When used in the form of bulk elemental sulfur, or micro- or nano-sulfur, applied both to the soil and to the canopy, the element undergoes a series of changes in its oxidation state, produced by various intermediaries that apparently act as biostimulants and promoters of stress tolerance. The final result is sulfate  $S^{+6}$ , which is the source of sulfur that all soil organisms assimilate and that plants absorb by their root cells. The changes in the oxidation states of sulfur  $S^0$  to  $S^{+6}$  depend on the action of specific groups of edaphic bacteria. In plant cells,  $S^{+6}$  sulfate is reduced to  $S^{-2}$  and incorporated into biological molecules.  $S^{-2}$  is also absorbed by stomata from  $H_2S$ , COS, and other atmospheric sources.  $S^{-2}$  is the precursor of inorganic polysulfides, organic polysulfanes, and  $H_2S$ , the action of which has been described in cell signaling and biostimulation in plants.  $S^{-2}$  is also the basis of essential biological molecules in signaling, metabolism, and stress tolerance, such as reactive sulfur species (RSS), SAM, glutathione, and phytochelatins. The present review describes the dynamics of sulfur in soil and plants, considering elemental sulfur as the starting point, and, as a final point, the sulfur accumulated as  $S^{-2}$  in biological structures. The factors that modify the behavior of the different components of the sulfur cycle in the soil-plant-atmosphere system, and how these influences the productivity, quality, and stress tolerance of crops, are described. The internal and external factors that influence the cellular production of  $S^{-2}$  and polysulfides vs. other S species are also described. The impact of elemental sulfur is compared with that of sulfates, in the context of proper soil management. The conclusion is that the use of elemental sulfur is recommended over that of sulfates, since it is beneficial for the soil microbiome, for productivity and nutritional quality of crops, and also allows the increased tolerance of plants to environmental stresses.

## Keywords:

plant nutrition; sulfate; sulfite; plant health and nutrition; nutraceuticals; polysulfanes; polysulfides; soil microbiome

## 1. Introduction

Sulfur is one of the most abundant elements on Earth and is an essential element for living beings, of which constitutes on average 1% of dry weight. In plants, S content varies strongly between species, ranging from 0.1 to 6% of dry weight (0.03 to 2 mmol  $g^{-1}$  dry weight) [1]. S belongs to the VIA group of the periodic system, where it is found together with O, Se, Te, and Po; naturally, S is a mixture of four isotopes,  $^{32}S$ ,  $^{33}S$ ,  $^{34}S$ , and  $^{36}S$ . The natural abundance of each is 95.1%, 0.74%, 4.2%, and 0.016%, respectively. Sulfur exists in oxidation states ranging from +6 to -2 (Table 1), with the most oxidized state

in the form of sulfate ( $\text{SO}_4^{2-}$ ), which is the chemical form that plants absorb from the soil to feed themselves with S [2].

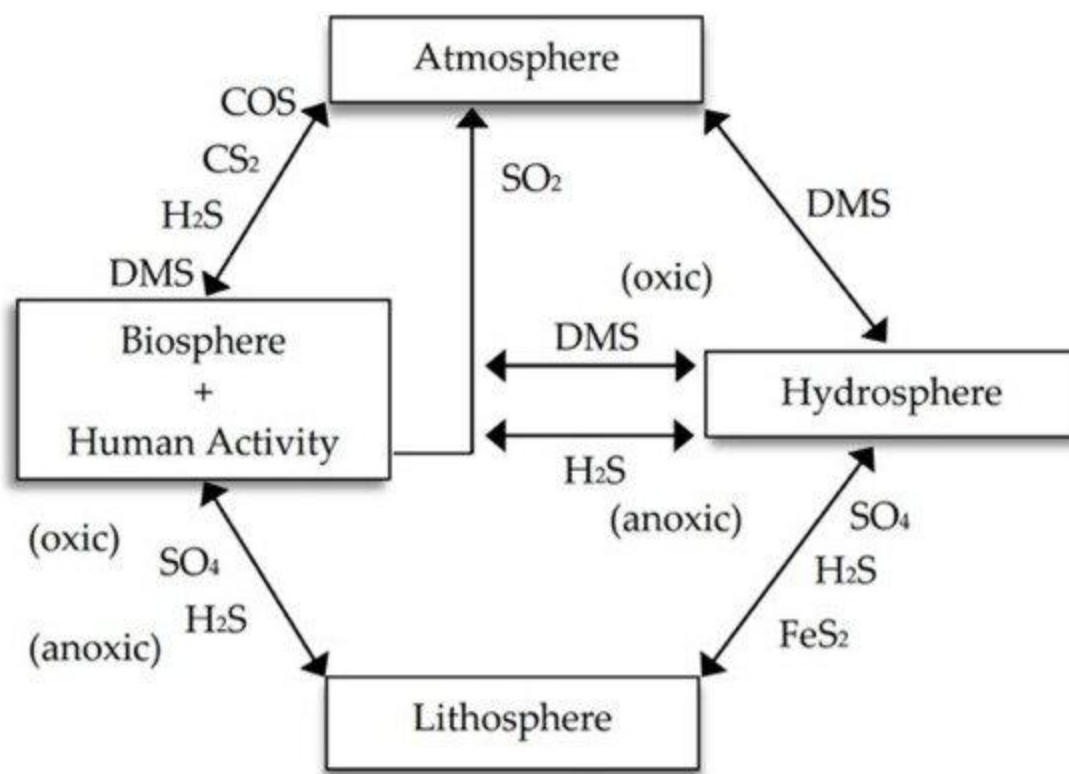
Biological molecules, which range from small molecules to proteins and other polymers, contain S in its more reduced states 0, -1, and -2. For example, it is known that approximately 40% of enzymes depend for their catalytic activity on the presence of sulfhydryl groups (-SH). These -SH groups participate in redox reactions, provide binding sites for toxic or physiologically important metals, and are related to the detoxification of various xenobiotics. It is also known that the tertiary and quaternary structure of many proteins is the result of the presence of disulfane bonds (-S-S-) formed by the oxidation of -SH groups of cysteine, a sulfur amino acid that, together with methionine, is a key factor in determining the nutritional value of plants, as well as a central element in the metabolism of S in all organisms [2].

For the above reasons, a close relationship between nitrogen and sulfur nutritional status has been found in plants [3,4]. Approximately 80% of nitrogen and sulfur incorporated in organic compounds of plants is found in proteins when both elements are in adequate proportions. The S/N balance of a plant, described by the organic S/N ratio, is in the range of 0.025 (legumes) to 0.032 (grasses) and is relatively constant from one species to another. Therefore, the amount of S required by a plant is strongly dependent on its N nutrition. The consequence is that the availability of S below the needs of the crops does not allow the adequate use of applied N [5].

Compounds as important as  $\beta$ -lactam antibiotics (penicillins, cephalosporins, and cephamycins) have an S atom derived from cysteine. The sulfur compound S-adenosyl-L-methionine (SAM) is the most crucial methylating agent known in all organisms; SAM-mediated transmethylation reactions are essential in the regulation of gene expression, the activity of various enzymes, the synthesis of compounds such as the osmolyte DMSP (dimethyl sulfoniopropionate) and DMS (dimethyl sulfide) gas, as well as in the production of antibiotics [2].

The Earth's S stores are located in the lithosphere, hydrosphere, atmosphere, and biosphere. Human activities result in the extraction of S from the lithosphere (burning of fossil fuels, mining of elemental S and metals) and biosphere (oxidation of organic matter from the soil and burning of biomass). Anthropogenic S is incorporated into the global cycle mainly in the form of  $\text{SO}_2$  emitted into the atmosphere [6].

Between the terrestrial and marine masses, there is a constant flow of S via the atmosphere through the gaseous forms of the element ( $\text{SO}_2$ , COS,  $\text{H}_2\text{S}$ , DMS, and  $\text{CS}_2$ ) [6] and aerosols (mainly  $\text{SO}_4^{2-}$  from the oxidation of sulfur gases, and <10% of organosulfates) [7], or by runoff from terrestrial to oceanic regions (Figure 1). The constant mobilization of S causes changes in the sulfur species that move from one terrestrial compartment to another. Under oxic conditions, the predominant inorganic form of S is  $\text{SO}_4^{2-}$ , resulting from atmospheric deposition or oxidation of reduced forms of S. In the soil, continuous land tillage that oxidizes soil organic matter and repeated extractions for crops cause the decrease of S stores; for this reason, the regular application of S with the fertilizers is recommended [8].



**Figure 1.** Simplified biogeochemical sulfur cycle. Human activities, fauna, vegetation, and soil microorganisms can be visualized as an interface (as source and sink) to accelerate the transfer of sulfur species between the lithosphere, atmosphere, and hydrosphere.

In soils, most S is found in organic forms; the inorganic forms are elemental sulfur ( $S^0$ ) or  $SO_4^{2-}$ , the latter can be found as gypsum or be adsorbed in the inorganic exchange matrix. The  $SO_4^{2-}$  adsorbed in the soil is in dynamic equilibrium with the soil solution, and the adsorption/desorption quotient inversely depends on the pH value of the soil and the cations present in the exchange matrix, showing higher affinity for  $Al^{3+} > Ca^{2+} > K^+$  [9,10].

In soil,  $SO_4^{2-}$  is subject to dissimilatory and assimilatory reduction. Dissimilatory reduction occurs when  $SO_4^{2-}$  is used as a final acceptor of electrons in the anaerobic metabolism of microorganisms, producing  $H_2S$  that is reoxidized in the presence of  $O_2$  or volatilized into the atmosphere. Assimilatory reduction is used by prokaryotes, algae, plants, and fungi for the biosynthesis of organic compounds, e.g., amino acids. Animals and protists cannot perform assimilatory reduction of  $SO_4^{2-}$ ; therefore, they depend on the organic sulfur compounds synthesized by other organisms [6]. In many crop species, sulfur is an element associated with nutritional quality and density of mineral nutrients, tolerance to stress, and the management of certain pests and pathogens [11,12,13].

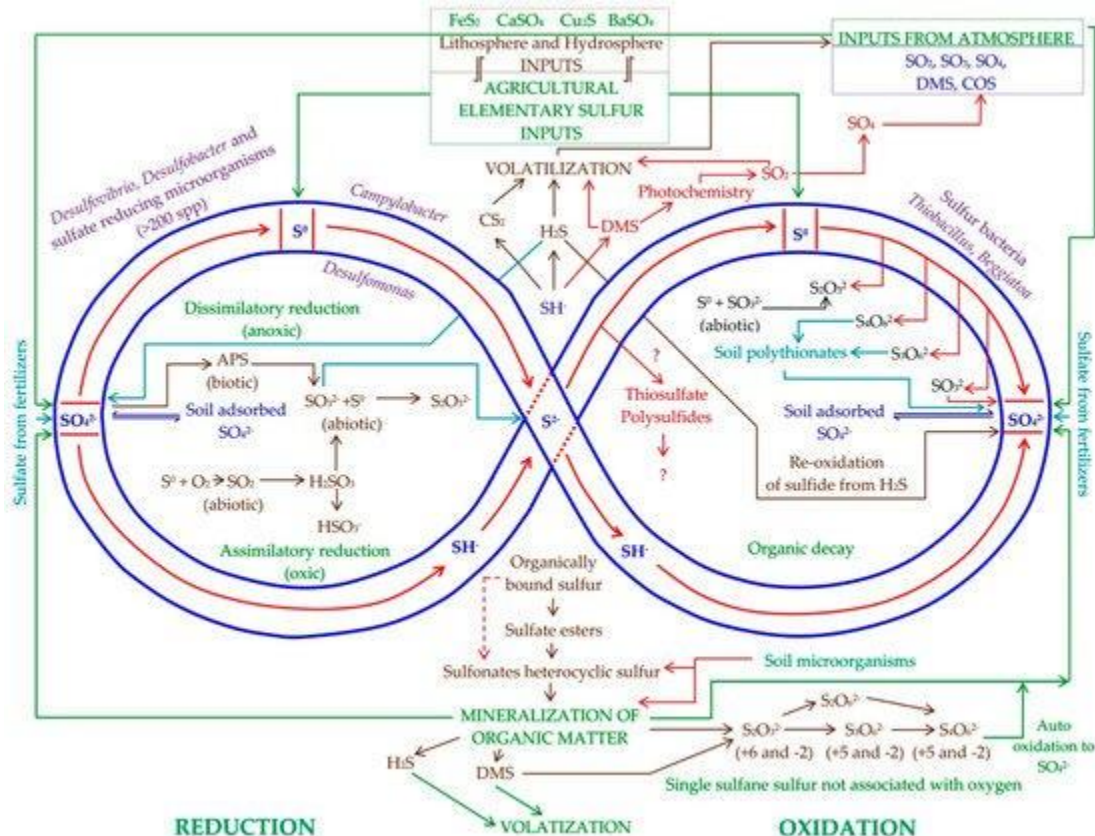
In agricultural soils,  $SO_4^{2-}$  used by crop plants comes mainly from the contribution of fertilizers with sulfates, such as ammonium sulfate, gypsum, potassium sulfate, magnesium sulfate, single superphosphate, ammonium phosphate sulfate, potassium magnesium sulfate, and sulfates of micronutrients [8,14], as well as the oxidation of  $S^0$ , and of  $S^{2-}$  contained in organic fertilizers. Another part of the S of crops is obtained from  $SO_4^{2-}$  and aerosols coming from precipitation, as well as the absorption by soil and plants of aerosols and gases such as  $H_2S$ , COS, and DMS. When S is added to the soil in the form of  $SO_4^{2-}$ , plants and aerobic prokaryotes absorb it and incorporate it into a reductive metabolism that produces sulfide ( $S^{2-}$ ). On the other hand, when S is supplied as  $S^0$  or in the form of organic fertilizers ( $S^{2-}$ ), it must be oxidized to  $SO_4^{2-}$  by the action of soil prokaryotes to be available to plants [2,6,8].

The aim of the present review is to describe the dynamics of sulfur in soil and plants, considering elemental sulfur as the starting point and, as a final point, the sulfur accumulated as  $S^{2-}$  in biomolecules and biological structures, transformed into myriad sulfur compounds and returned to atmosphere and

hydrosphere as  $\text{H}_2\text{S}$  and other gaseous molecules. The factors that modify the behavior of the different components of the sulfur flow in the soil–plant–atmosphere system are described, along with how these influences the productivity, quality and stress tolerance of crops.

## 2. Transformations of Elemental Sulfur in Soil

The S available for plants in agricultural ecosystems is in dynamic storage (Figure 2). It comes from gaseous forms and aerosols of S from the atmosphere, from dissolved S (mostly  $\text{SO}_4^{2-}$ ) in rain and snow precipitation, and from  $\text{SO}_4^{2-}$ , which is obtained from the oxidation of S of soil organic matter and  $\text{S}^0$ . Sulfates can be fixed in the soil exchange matrix or leached to the subsoil [10]. In arid regions,  $\text{SO}_4^{2-}$  can be stored in large quantities as gypsum in the subsoil, but in areas with higher water availability, leached  $\text{SO}_4^{2-}$  is mobilized to lower horizons and to the subsoil [15].



**Figure 2.** Schematic representation of the flow of sulfur in soil. APS = adenosine 5'-phosphosulfate. Oxidation states of sulfur in the different molecules are:  $\text{SO}_4^{2-}$  (+6);  $\text{S}_2\text{O}_6^{2-}$  (+5 and -2);  $\text{S}_4\text{O}_6^{2-}$  (+5 and -2);  $\text{S}_3\text{O}_6^{2-}$  (+5 and -2);  $\text{SO}_3^{2-}$  (+4);  $\text{SO}_2$  (+4);  $\text{S}_2\text{O}_3^{2-}$  (+6 and -2);  $\text{COS}$  (+2);  $\text{S}^0$  (0);  $\text{SH}^-$  (-2);  $\text{S}^{2-}$  (-2);  $\text{DMS}$  (-2);  $\text{CS}_2$  (-2).

In the anoxic zones of the soil,  $\text{S}^0$  and  $\text{SO}_4^{2-}$  are transformed to  $\text{H}_2\text{S}$  that is volatilized or is reoxidized to  $\text{S}^0$  and sulfate in the oxic zone. Plants and microorganisms take the  $\text{SO}_4^{2-}$  and reduce it to  $\text{S}^{2-}$  to incorporate it into a huge variety of organic compounds. Subsequently, these same plants and microorganisms transform a part of the sulfur to  $\text{H}_2\text{S}$ ,  $\text{DMS}$ , and  $\text{CS}_2$  [8,16]. The above volatile molecules have been associated with detoxification metabolism, stress tolerance, and signaling in plants and prokaryotes [17,18]. As with iodine [19], soil organic matter can transform the S to volatile forms by means of abiotic reactions, but the rate of transformation is very low in comparison with biotic metabolism of S [16].

Since there are several access ways by which S can enter the agricultural ecosystem, it is not possible to mark a specific starting point. Therefore, arbitrarily, the assumption of an application of  $\text{S}^0$  to the soil is taken, and the transformations that this material experiences up to  $\text{SO}_4^{2-}$  are described. Once in the form of  $\text{SO}_4^{2-}$ , it is assimilated into plant cells in the form of myriad organic compounds. The final



part of the flow of S from soil to plants ends with the production of volatile compounds by plant cells, or in the transformation of the S contained in plant waste (Figure 2).

S atoms tend to avoid double bonds, therefore, in the  $S^0$ , instead of forming molecules of  $S_2$  ( $S=S$ ) the S atoms are grouped in the form of cyclic allotropes (cyclosulfur) or as long chains  $S_n$  (catena sulfur) [20]. The  $S^0$  used to apply to soil consists mainly of molecules of  $S_8$  (cycloocta-S) that are grouped, forming polymers of variable size;  $S_8$  is the most stable form from a thermodynamic point of view.  $S_8$  is a very electrophilic Lewis acid, so it reacts with nucleophilic anions or Lewis bases such as  $OH^-$ , sulfides ( $S^{2-}$ ), thiols ( $R-SH$ ), thiolates ( $RS^-$ ),  $I^-$ ,  $CN^-$ , and  $SO_3^{2-}$  [21,22].

$S^0$  is applied to the soil or substrate in quantities ranging from 20 to 250 kg ha<sup>-1</sup> yr<sup>-1</sup>, the last figure being equivalent to 200 mg  $S^0$  kg<sup>-1</sup> soil. Once in the soil or substrate,  $S^0$  begins to transform into other chemical forms, mainly through biotic processes, and, to a lesser extent, by abiotic processes. The transformation rate is inversely proportional to the particle size and directly proportional to the temperature ( $Q_{10} = 4.0$ ), humidity availability, and abundance of edaphic microorganisms [8,23,24].

Any factor that decreases bacterial activity, such as temperatures <10 °C or >40 °C and lack of humidity in the soil, will reduce the transformation of S. Flooded or compact soils will have anoxic conditions that induce high rates of conversion of  $S^0$  and  $SO_4^{2-}$  into gaseous forms of sulfur [8,24]. The metabolism of S in soils can modify other processes, as in rice paddies, where the use of gypsum amendment has been shown to decrease greenhouse methane emissions [25]. In alkaline soils, it has been observed that the use of  $S^0$  induces acidification (by  $H_2SO_4$ ), which increases the bioavailability of elements such as P [26].

When it is desired that  $S^0$  produces  $SO_4^{2-}$  rapidly available for crops, an  $S^0$  source with a small particle size (<150 µm or 100 mesh) should be chosen. Contrarily, if a long-term impact (two or more consecutive crops) is sought, it is desirable to use  $S^0$  sources with a larger particle diameter, or even granular forms such as  $S^0$  prills or  $S^0$ -fortified N-P-K and DAP fertilizers [27,28]. At a temperature of 14 °C, it was found that, in 51 weeks, 51% of  $S^0$  with particle diameter 41 µm (300 mesh) was oxidized, compared to 18% of  $S^0$  with 125 µm (120 mesh). In soils with low temperatures,  $S^0$  sized 41 µm will oxidize at a rate equivalent to  $S^0$  sized 125 µm in soils with higher temperatures [23]. In another experiment, applying 50 kg ha<sup>-1</sup> of  $S^0$ , it was found that 80–90% of  $S^0$  with particles <150 µm was oxidized over a period of 340 days [29].

On the other hand, it has been found that repeated applications of  $S^0$  to soil increase the population and the activity of oxidizing bacteria of  $S^0$  [24]. Accompanying the increase in  $S^0$  oxidant bacteria was a reduction in the number of fungi and protists, while bacterial and actinomycete populations remained stable [30]. Other authors reported a decrease in biomass and bacterial metabolism by applying  $S^0$  annually for five years [31].

When  $S^0$  is in micronized form (<177 µm, <80 mesh) it is used for the control of mites and some fungi [32,33]. The reactivity of micronized  $S^0$  is a consequence of the high quotient surface/volume of the particles, estimated to be 1300 to 1940 cm<sup>2</sup> g<sup>-1</sup> for  $S^0$  of 125 and 41 µm, respectively [23]. Micronized  $S^0$  can be applied through the foliar route or even by using pressurized irrigation systems to incorporate it into the soil [34,35]. When applied by irrigation system, the problems associated with the application of micronized  $S^0$  (because it is a flammable and irritant material) by dusting machines are reduced [24].

The use of S nanoparticles for the control of pathogens in plants has also been described [36,37]. Taking into account the high value of the surface/volume ratio of S nanoparticles, furthermore being a source of S for rapid assimilation by plants and microorganisms, it is possible that they function as biostimulants [38], and that they provide highly reactive  $S^0$  that works as a tolerance-inducing factor against pathogenic fungi [32,39].

To be available for plants,  $S^0$  applied to the soil or substrate must be oxidized to  $SO_4^{2-}$ . The change in the oxidation state of sulfur from 0 to +6 allows reduction equivalents to be obtained ( $8H^+ + 6e^-$ ). The oxidation is carried out by most soil microorganisms, highlighting *Thiobacillus*, *Beggiatoa*, *Desulfomicrobium*, and *Desulfovibrio*, as well as other heterotrophic aerobics S-oxidizing bacteria such as *Bacillus*, *Pseudomonas*, and *Arthrobacter* [2,40]. Two metabolic pathways have been described that allow the oxidation of inorganic S to  $SO_4^{2-}$ : the Kelly–Friedrich pathway, which does not involve the production of intermediates such as polythionates, and the Kelly–Trudinger pathway, which includes as an intermediate output tetrathionate ( $S_4O_6^{2-}$ ) and other polythionates [16]. The existence of two different routes and the large number of taxa that carry out the oxidation of  $S^0$  allow a high redundancy, and capacity to tolerate extensive changes in pH and salinity in soils [41,42].

In [Figure 2](#), the oxidation activity from  $S^0$  to  $SO_4^{2-}$  is presented on the right side, and shows the Kelly–Trudinger pathway with the production of polythionates such as  $S_4O_6^{2-}$  and  $S_3O_6^{2-}$  (as well as  $S_2O_3^{2-}$  and  $SO_3^{2-}$ ), which serve as a source of reducing potential and possibly act as inducers of stress tolerance in plants, perhaps by containing a single sulfane sulfur not associated with oxygen [16]. In this regard, Li et al. [43] described polythionates as agents with antibiotic action, the efficacy of which is variable according to the pH. Additionally, the abiotic oxidation of  $S^{2-}$  in the presence of  $S^0$  produces polysulfides [44], which have been described as agents associated with stress tolerance in animal cells [45]. Polysulfides possibly fulfill a similar stress-protection function in plants [46]. It is possible that the presence of  $S^0$  and  $S^{2-}$  in polysulfides [44] explains their ability to induce stress tolerance. The production of polythionates and polysulfides represents an additional advantage of the use of  $S^0$  as a source of sulfur for crops.

Under anoxic conditions,  $S^0$  is produced as part of the dissimilatory reduction of  $SO_4^{2-}$ . Later, the  $S^0$  can be assimilated into  $S^{2-}$  that will be part of the biomolecules, or it will be volatilized in the case of excess S (see the central section of [Figure 2](#)). At the left and right ends of [Figure 2](#), in the central part, the  $SO_4^{2-}$  from fertilizers, precipitation, and mineralization of organic matter is represented. A portion of this  $SO_4^{2-}$  forms a soil adsorbed sulfate storage, which will be in dynamic equilibrium with  $SO_4^{2-}$  dissolved in the soil solution. Under oxic conditions,  $SO_4^{2-}$  is assimilated in  $S^{2-}$  by assimilatory reduction and then transformed back into  $SO_4^{2-}$  during the organic decay and mineralization of organic matter [42].

As part of the processes of organic decay, mineralization of organic matter, and sulfate reduction, both the soil, through abiotic reactions, and micro-organisms and plants can be source or sink of volatile forms of S, such as  $H_2S$ , DMS, COS,  $CS_2$ , and  $SO_2$  ([Figure 2](#)). Generally, under anoxic conditions, the oxidized forms of S are reduced by the soil microbiome to  $H_2S$ ,  $CS_2$ , COS, DMDS, methyl mercaptan, and COS [8]. These gaseous molecules are believed to be part of a mechanism of dissipation of excess S, although participation in other processes is not ruled out [16,47].

In terms of reductive and oxidative microbial reactions, the most abundant forms of sulfur in the soil and the edaphic microbiome are  $S^{2-}$ , R-SH, RSSH, polysulfides ( $RS_n^{2-}$ ),  $S_2O_3^{2-}$ ,  $SO_3^{2-}$ ,  $SO_4^{2-}$  and polythionates [16].  $SO_4^{2-}$  applied as fertilizer or obtained through the processes described above is the form of S that plants assimilate through their roots [40].

At best growth conditions, a plant's sulfur requirement ranges from 2 to 10  $\mu\text{mol g}^{-1}$  plant fresh weight day $^{-1}$  [1]. As the flow of S is a dynamic process where the ecosystem receives S from the atmosphere, precipitation, subsoil water, and fertilizers, and loses S through the process of volatilization of S by soil and plants and by leaching, it is difficult to estimate the actual amount of S that a plant surface absorbs, assimilates, leaches, and volatilizes. As an exercise, let us suppose a single sampling point for a field of maize, for example before the harvest. In one hectare, there may be 78,000 kg of fresh weight  $\text{ha}^{-1}$ , which would be equivalent to 25 kg of sulfur contained in the plants. However, the 25 kg  $\text{ha}^{-1}$  accumulated in the plant tissues at that specific sampling time does not include the S volatilized by the plant itself, or that leached, assimilated, or volatilized in the soil and by the microorganisms.

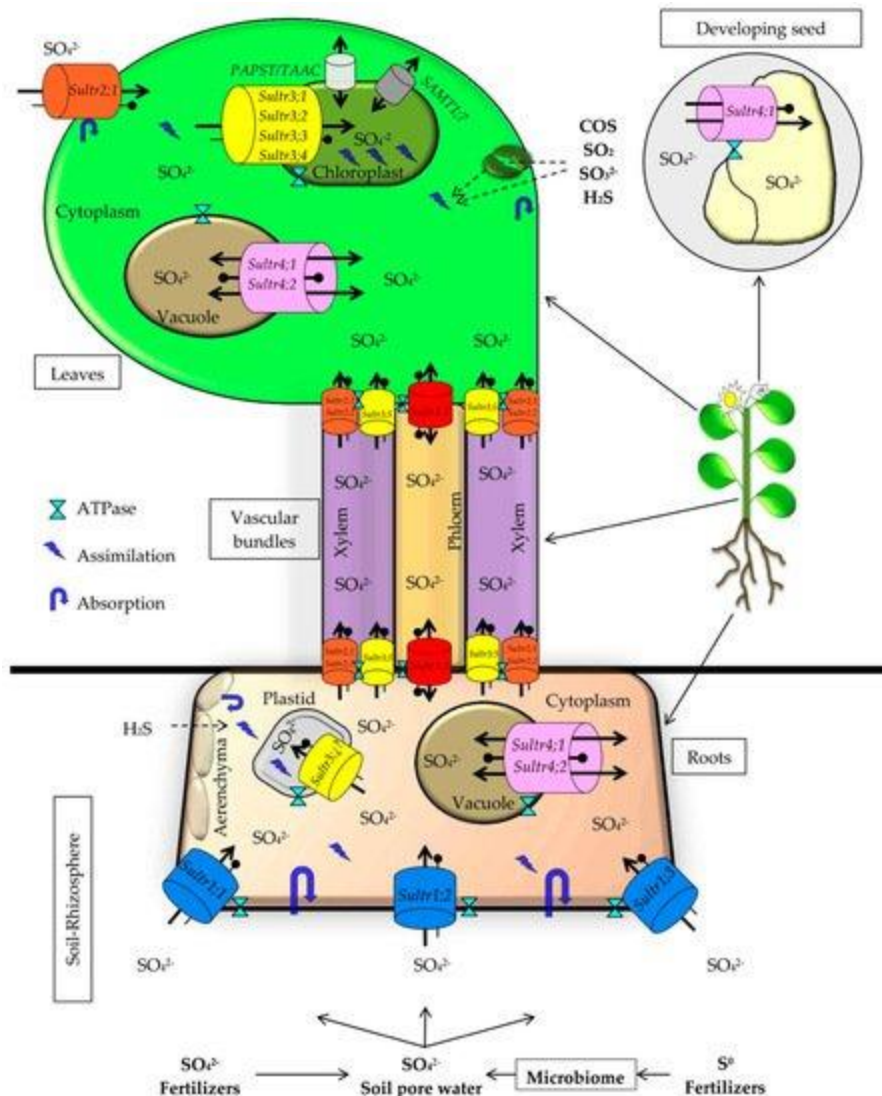
The point to highlight with the data of the previous paragraph is that the S of the soil is in constant exchange and extraction by the crops, atmosphere, and soil water. Therefore, a continuous supply of S is required, which is recommended to be applied in the form of  $S^0$  (40–60 kg  $\text{ha}^{-1}$ ) every one or two years, to maintain the edaphic store.

### 3. Absorption and Assimilation of Sulfur in Plants

#### 3.1. Sulfur Absorption and Transport

The absorption of S from atmospheric sources such as COS,  $SO_2$ , DMS, and  $H_2S$ , can represent a valuable contribution of sulfur for many plants. However, most of the S taken by the plants comes from  $SO_4^{2-}$  dissolved in the soil solution [40,47].

The  $SO_4^{2-}$  dissolved in the soil solution is absorbed by  $H^+$ /sulfate cotransporters called SULTRs ([Figure 3](#)). Plant SULTRs are encoded by a multigene family. SULTRs include high-affinity transport proteins (HAST), low-affinity transport proteins (LAST), vacuole transporters, and plastid membranes and endosymbionts transporters [48,49]. The level of  $SO_4^{2-}$  in the soil solution that induces high-affinity transporters is  $<10 \text{ mg L}^{-1}$  (0.1 mM) [1]. In the soil solution of agricultural areas under strong fertilization management, values of 40–200  $\text{mg L}^{-1}$   $SO_4^{2-}$  are found, while for non-agricultural fertile soils, concentrations of  $SO_4^{2-}$  of 4.5–40.5  $\text{mg L}^{-1}$  were reported in the soil solution [50]. The assimilation of S shows a high degree of control and coordination with the assimilation of C and N, and in root transporters there seems to be a relevant regulatory site [1,49].



**Figure 3.** Schematic representation of the processes of absorption, transport, and storage of sulfate.

HAST Sultr1;1, Sultr1;2, and Sultr1;3 facilitate the absorption of  $\text{SO}_4^{2-}$  in the root. The HAST of the epidermis and the cortex is accompanied by the LAST Sultr2;1, Sultr2;2, and Sultr3;5, with which they act synergistically. HAST are very abundant in the epidermis and cortex of the root, while LAST proliferate in the parenchyma adjacent to the xylem and phloem [49,51]. The  $\text{SO}_4^{2-}$  absorbed is stored in the vacuoles thanks to the co-transporters Sultr4;1 and Sultr4;2 [49,51,52], or it is distributed to the rest of the plant, depending on the sink tissues' demand. Translocation from the root to the stems and leaves through the xylem occurs through Sultr1;3, Sultr2;1, Sultr2;2, and Sultr3;5 [49,51,53,54,55]. In most species, the absorbed sulfate is assimilated largely into proteins and other biomolecules. However, in others, such as *Brassica oleracea* seedlings, it is possible to observe high amounts of  $\text{SO}_4^{2-}$  in the plant tissues [1].

The discharge of  $\text{SO}_4^{2-}$  from the xylem to the mesophyll cells of the leaf is mediated by HAST and LAST (Sultr1;3, Sultr2;1, Sultr2;2, and Sultr3;5). As occurs in the root, a part of the sulfate is stored in the vacuoles of stems and leaves by Sultr4;1, and Sultr4;2 [49,51,52], while a part is taken to the chloroplasts by the co-transporters Sultr3;1, Sultr3;2, Sultr3;3, and Sultr3;4 [49,51,55,56], where it will be reduced to  $\text{S}^{2-}$  to be assimilated into biological molecules [56]. According to the plant's needs, the sulfate stored in the vacuoles can be re-mobilized through the co-transporters Sultr4;1 and Sultr4;2 [51,52].

The capacity of sulfur absorption, assimilation, and volatilization responds to the nutritional status of the plant. In turn, the sulfur nutritional status of the latter depends on the growth rate and the interaction with the C and N levels of the plants. Biomolecules synthesized from C and N assimilated during

photosynthesis are the primary sink for S, and probably constitute part of the signals that regulate  $\text{SO}_4^{2-}$  absorption, transport, and assimilation [1]. It is assumed that an excess of S will trigger a higher accumulation of  $\text{SO}_4^{2-}$  in the vacuoles, as well as an increase in the synthesis of volatile forms of S, such as  $\text{H}_2\text{S}$  [16,47]. On the other hand, under S deprivation there will be a significant increase in the expression of sulfate transporters, which will increase the absorption and assimilation of  $\text{SO}_4^{2-}$  [1].

Other environmental factors, which possibly modify the absorption of  $\text{SO}_4^{2-}$ , also regulate the relative abundance of HAST and LAST. Among these factors are salinity, drought, and high temperature in maize [57], drought and salinity in *Arabidopsis* and *Medicago truncatula* [58], and heavy metals such as Cd in sorghum [59].

Considering now the absorption of gas molecules of S, atmospheric  $\text{SO}_2$  can be absorbed through the stomata. In the water film of the substomatal chamber, it is transformed into  $\text{HSO}_3^-$  and  $\text{SO}_3^{2-}$ , which is incorporated into the sulfur reduction pathway to be reduced to  $\text{S}^{2-}$ . Another alternative is that  $\text{SO}_3^{2-}$  can be oxidized extra- and intracellularly to  $\text{SO}_4^{2-}$  by peroxidases, or non-enzymatically by  $\text{O}_2^-$  radicals or metal ions. This new  $\text{SO}_4^{2-}$  is again incorporated into sulfur reduction pathway or transferred into the vacuole. A high level of  $\text{SO}_4^{2-}$  is typical in plants exposed to  $\text{SO}_2$  [1,60,61].

Similarly, by diffusion,  $\text{H}_2\text{S}$  from the soil, vegetation, or atmosphere can be absorbed directly or dissolved in air  $\text{H}_2\text{O}$  aerosols, through stomata [62,63,64]. In the mesophyll,  $\text{H}_2\text{S}$  is assimilated by O-acetyl-serine (thiol)lyase for the biosynthesis of cysteine [60]. The exogenous application of  $\text{H}_2\text{S}$  has been described as a factor that increases tolerance to water deficit directly through a higher content of cysteine, and indirectly through the synthesis of metabolites such as proline and glycine betaine, and upregulating antioxidant enzymes [65].

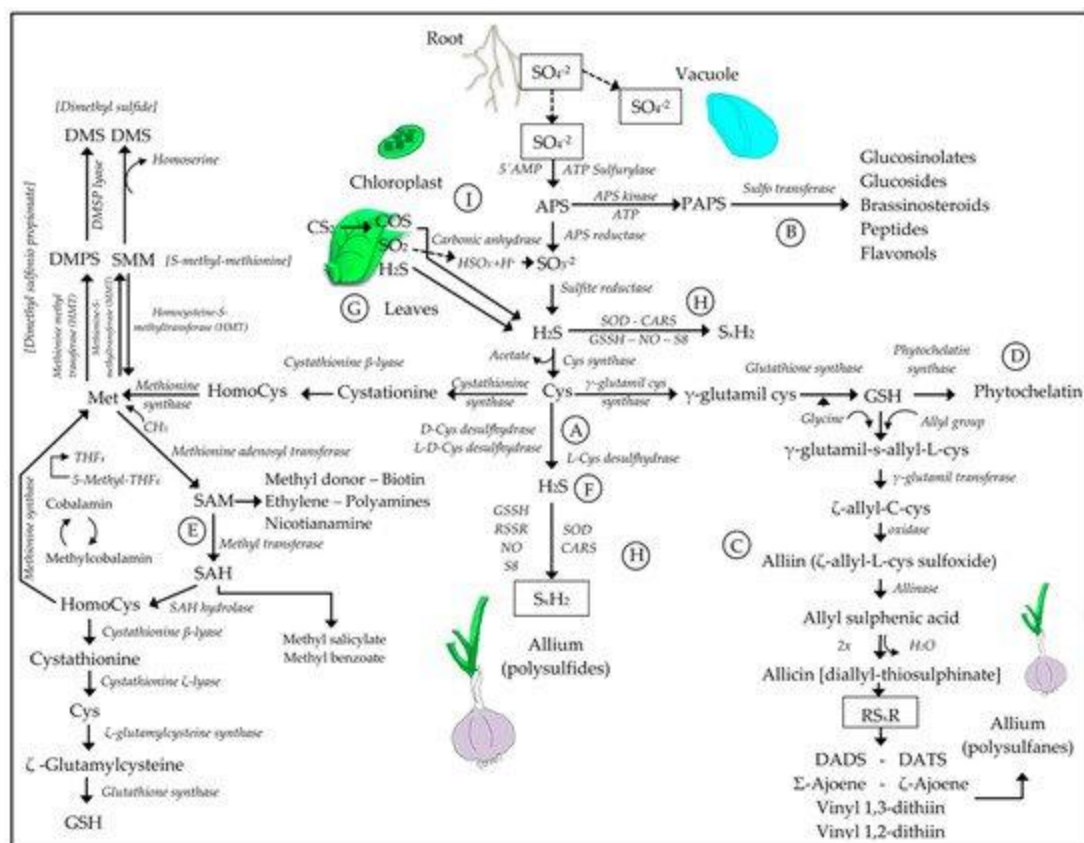
Additionally, a synergistic interplay between nitric oxide (NO) and  $\text{H}_2\text{S}$  signaling has been shown during stress events [66]. In anoxic soils, the aerenchyma of the root intervenes in the fixation of the  $\text{H}_2\text{S}$  produced by microorganisms [67], so that the  $\text{H}_2\text{S}$  of the soil can also trigger adaptive responses to stress in the radical tissues. On the other hand, the absorption and assimilation of  $\text{SO}_2$  and  $\text{H}_2\text{S}$  in the leaves is a factor that modifies the nutritional status of the sulfur in the plant and decreases the activity of the  $\text{SO}_4^{2-}$  transporters in the root [60].

In the case of COS, this gas is produced by biotic synthesis or by the oxidation of  $\text{CS}_2$  and DMDS. In plants, the  $\text{SCN}^-$  obtained from glucosinolate metabolism is hydrolyzed to COS and  $\text{NH}_3$ . COS is the most stable and abundant gas form of S in the atmosphere (0.5 ppb). In the atmosphere, COS can be oxidized to sulfate to form aerosols, or be absorbed by plants and microorganisms that transform it into  $\text{CO}_2$  and  $\text{H}_2\text{S}$  through carbonic anhydrase, RUBISCO, nitrogenase, and other metalloenzymes. In plants, COS can also be reduced to  $\text{CS}_2$ , which is released back into the atmosphere [18]. The  $\text{SO}_4^{2-}$  from the aerosols can be absorbed in the leaves by means of the co-transporter Sultr2;1 [62], located in the cells of the mesophyll.

### 3.2. Sulfur Assimilation and the Synthesis of $\text{H}_2\text{S}$

Once  $\text{SO}_4^{2-}$  is available in the cells from root absorption and transport, or by absorption of gaseous forms of S by the stomata, (Figure 4G,I) it is used to assimilate S into biomolecules [68,69,70]. The first step is to activate  $\text{SO}_4^{2-}$  by means of the enzyme ATP sulfurylase. The resulting compound, adenosine 5'-phosphosulfate (APS), is used as a bifurcation between two assimilation pathways, primary and secondary. In the primary pathway, the APS is reduced by APS reductase to  $\text{SO}_3^{2-}$ , which in turn is reduced to  $\text{S}^{2-}/\text{H}_2\text{S}$ , which is assimilated into the amino acid cysteine (Figure 4A) [68,71]. Cysteine is the first product of the primary assimilation pathway of  $\text{SO}_4^{2-}$ , and is used for the synthesis of methionine and proteins, or as a donor of  $\text{S}^{2-}$  for the synthesis of a large number of metabolites, such as  $\text{H}_2\text{S}$  (Figure 4F) [72], SAM, GSH, phytochelatins (Figure 4D,E) [48,68], SAM [68,73], polysulfides [72], and polysulfanes [74,75] (Figure 4C,H). In the secondary pathway (Figure 4B),  $\text{SO}_4^{2-}$  is phosphorylated by APS kinase to 3'-phosphoadenosine 5'-phosphosulfate (PAPS) [48,68]. PAPS is the active sulfate donor for a variety of sulfation reactions in secondary metabolism. The reaction of sulfation is catalyzed by sulfotransferases, located mainly in the cytoplasm. To date, a large number of sulfated secondary compounds have been found to be involved in growth and stress signaling, and in the detoxification of environmental toxins [76,77].





**Figure 4.** Schematic representation of the primary and secondary pathways of sulfur assimilation. In the primary assimilation pathway (A), APS is reduced to  $\text{SO}_3$  and subsequently to  $\text{S}^{2-}/\text{H}_2\text{S}$ , which are assimilated to form the amino acid cysteine [68,71]. In the secondary pathway (B),  $\text{SO}_4$  is phosphorylated and converted to 3'-phosphoadenosine 5'-phosphosulfate (PAPS) [48,68]. Cysteine is a central point for the synthesis of methionine or the production of polysulfanes (C), polysulfides (H), phytochelatins (D), SAM (E), and  $\text{H}_2\text{S}$  (F) [48,68,72,73,74,75]. The absorption of sulfur in its gaseous forms is carried out by the stomatal route, directly incorporated into the primary pathway ( $\text{SO}_2$  and  $\text{H}_2\text{S}$ ) (G) [68], or through the action of carbonic anhydrase (COS) (I) [69,70].

As occurs in the metabolism of S in the soil (Figure 2), the great diversity of oxidation states of S allows the construction of a sophisticated and rich network of functional sulfur molecules for cell metabolism and signaling of plants. As in the soil, plants can exchange sulfur from the plant to the atmosphere and vice versa in the form of  $\text{H}_2\text{S}$ , COS, and  $\text{CS}_2$ , to incorporate the S in the assimilation pathways shown in Figure 4.

In recent years, both  $\text{H}_2\text{S}$  and several of the derivative compounds (polysulfides and polysulfanes) or donors of this molecule have aroused great interest for their participation as oxidative stress reducers, in cellular signaling, and as post-translational modifiers. Because of its lipophilic nature,  $\text{H}_2\text{S}$  is biologically reactive, since it can rapidly cross the membranes of cells without the intervention of channels. A possible  $\text{H}_2\text{S}$  signaling mechanism is the formation of persulfides or hydrosulfides (RSSH) from the protein cysteine residues. It is assumed that  $\text{H}_2\text{S}$  interacts in this way, with a great diversity of proteins such as channels, transcription factors, and enzymes [78,79].

$\text{H}_2\text{S}$  autooxidises in the presence of  $\text{O}_2$ , forming polysulfanes,  $\text{SO}_3^{2-}$ ,  $\text{S}_2\text{O}_3^{2-}$ , and  $\text{SO}_4^{2-}$  [70]; additionally,  $\text{H}_2\text{S}$  is also a precursor of biological polysulfides [72]. Polysulfanes, polysulfides (with  $\text{Sn} > 2$ ), and RSSH contain  $\text{S}^0$  atoms, which allows a diversity of oxidation states between the sulfur atoms and allows the molecules a dual character as oxidants and reducers. This diversity probably contributes to a multifunctionality character of the signaling of  $\text{H}_2\text{S}$  and its derived compounds. Although  $\text{H}_2\text{S}$  has a reactivity comparable to that of GSH against  $\text{H}_2\text{O}_2$  and free radicals, it is believed that its value as a cellular antioxidant is limited because of its low concentration in vivo [78,79].

H<sub>2</sub>S donor compounds have been explored in the agricultural field for their possible applications in improving the productivity and quality of crops. It has been found that H<sub>2</sub>S mediates in signaling and in the increase in tolerance to different stresses such as heavy metals (Cd, Cr, Cu, Al, As), salinity, high temperature, and water deficit [17,65,80,81,82,83,84,85]. Additionally, H<sub>2</sub>S and reactive sulfur species (RSS) interact with other relevant signaling molecules such as reactive oxygen (ROS) and reactive nitrogen (RNS) species [66,86], so the set of reactive chemical species could form a cellular network of redox signals [87]. These facts emphasize on the one hand the importance of adequate crop nutrition with S and, on the other hand, they highlight the advantages of the use of S<sup>0</sup> applied to the soil and by dusting machine [33], because, presumably, S<sup>0</sup> is a source of RSS as polysulfanes and polysulfides.

In addition to its relevance as a cell signaling and tolerance inducer, H<sub>2</sub>S is a source of RSS, a group of molecules of great biological importance that includes polysulfides and polysulfanes, SO<sub>2</sub>, S<sub>2</sub>O<sub>3</sub><sup>2-</sup>, allicin, diallyl disulfide (DADS), and diallyl trisulfane (DATS), shown in Figure 4C,H. RSS can also be formed by the oxidation of thiols (e.g., GSH oxidized by H<sub>2</sub>O<sub>2</sub>). RSS are sulfur species capable of initiating oxidation reactions by nucleophilic substitutions, and can be non-radical or radical, as the thiyl radical RS<sup>\*</sup>. RSS are not inactivated by antioxidants such as vitamins C and E or NADPH; for the above, GSH is required [88].

Polysulfides are inorganic RSS of the general formula RS<sub>n</sub><sup>2-</sup> ( $n > 2$ ) such as H<sub>2</sub>S<sub>2</sub><sup>2-</sup>, H<sub>2</sub>S<sub>3</sub><sup>2-</sup>, H<sub>2</sub>S<sub>4</sub><sup>2-</sup>, and H<sub>2</sub>S<sub>5</sub><sup>2-</sup>. Polysulfides are produced metabolically by enzymatic catalysis, by partial oxidation of S<sup>2-</sup> of H<sub>2</sub>S to produce H<sub>2</sub>S<sub>n</sub>, by reduction of H<sub>2</sub>S in the presence of polysulfanes, or by the reduction of polysulfanes in the presence of GSH. Polysulfides, depending on the molecule with which they interact, can behave as oxidants or reducers. It is considered that H<sub>2</sub>S<sub>n</sub> polysulfides could be part of the signaling network and antioxidant impact currently attributed entirely to H<sub>2</sub>S [72]. The IUPAC [89] defines polysulfides as compounds R-[S]<sub>n</sub>-R, with a chain of S atoms  $n \geq 2$  and R  $\neq$  H, however, in this manuscript we utilized the definition of Kharmas et al. [72].

In soil, the abiotic synthesis of polysulfides occurs through the reaction between S<sup>0</sup> and S<sup>2-</sup>, the oxidation of H<sub>2</sub>S by O<sub>2</sub>, H<sub>2</sub>O<sub>2</sub>, and possibly by iron oxides. Polysulfides are also produced by the bacterial oxidation of S<sup>2-</sup> and constitute an essential substrate for both aerobic and anaerobic microbial metabolism, for example during S<sup>0</sup> metabolism. In fact, it has been proposed that polysulfides and polysulfanes represent a critical part of the sulfur flux in ecosystems [44].

Polysulfanes are organic RSS with the general formula RS<sub>n</sub>H (R  $\neq$  H,  $n \geq 2$ ). Polysulfanes contain S<sup>2-</sup> and are very reactive with proteins and enzymes that contain cysteine; this characteristic possibly turns them into very versatile signaling agents [90].

The long-chain polysulfanes are characteristic molecules of the *Allium* species, with diallyl sulfanes in garlic and dipropyl sulfanes in onions [91,92]. The molecules frequently found in these species are allin (S-allyl-L-cysteine sulfoxide), diallyl trisulfane (DATS), diallyl tetrasulfane (DATTS), diallyl-pentasulfane (DAPS) or diallyl-hexasulfane, dimethyl-pentasulfane (DMPES), dipropyltrisulfane (DPTS), and dipropyl tetrasulfane (DPTTS), among others [90,91].

The compounds found in plants of the *Allium* genus have traditionally been used as plant protection products for both humans and crops [12]. This characteristic is believed to be related to the presence of polysulfides and polysulfanes in large quantities, 1.4% of fresh weight [93]. The reduced forms of polysulfanes can mitigate the impact of ROS such as superoxide and bind metal ions, decreasing oxidative stress in proteins and cell membranes [90]. The bioactivity of polysulfanes is enhanced by increasing the number of S atoms in the central functional groups of the molecule [94]. Pluth et al. [95] present in their review a large number of natural products that provide polysulfide-, polysulfane-, and H<sub>2</sub>S-releasing moieties.

The IUPAC [89] defines polysulfanes as having a chain  $> 2$  of unbranched S atoms terminating in H:HS<sub>n</sub>H, however, this manuscript has utilized the definition of Kharmas et al. [72].

#### 4. Conclusions

The use of elemental sulfur (S<sup>0</sup>) as a sulfur source for plants has several advantages over the use of sulfate fertilizers. The long residence time in the soil, the activation of the soil microbiome, and the transformation of S<sup>0</sup> into volatile sulfur species and reactive sulfur species that will induce higher tolerance to stress in plants were mentioned.

The process followed by sulfur was described, beginning with the application of S<sup>0</sup> to the soil or foliage until it is transformed into sulfate, which is incorporated into the metabolism of sulfur in the plant

through the primary and secondary pathways. Again, the application of  $S^0$  is associated with a great diversity of biomolecules that have a beneficial impact for plants, compared to the use of sulfur as sulfate.

The primary sulfate assimilation pathway was described as a process that gives rise to a great diversity of sulfur compounds, including  $H_2S$ , polysulfides, and polysulfanes, which increase the nutritional quality of plants and increase tolerance to biotic and abiotic stresses.

The sulfur nutrition of plants, especially using  $S^0$  to cover all or part of the sulfur needs of both the soil and plants, should be explored with a higher intensity as a sustainable technique for the management and care of crops.

### Author Contributions

Conceptualization, A.B.-M.; All authors were responsible for processing and organizing information; S.G.-M., A.J.-M., L.O.F.L., and A.B.-M. were responsible for manuscript drafting; All authors read and approved the final manuscript.

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### Conflicts of Interest

The authors declare no conflict of interest.

### References

1. De Kok, L.J.; Castro, A.; Durenkamp, M.; Koralewska, A.; Posthumus, F.S.; Stuiver, C.E.E.; Yang, L.; Stulen, I. Pathways of plant sulfur uptake and metabolism—An overview. In Proceedings of the 1st Sino-German Workshop on Aspects of Sulfur Nutrition of Plants, Shenyang, China, 23–27 May 2004; Schnug, E., de Kok, L.J., Eds.; FAL Agricultural Research: Braunschweig, Germany, 2005; pp. 5–13, ISBN 3-86576-007-4. [\[Google Scholar\]](#)
2. Huxtable, R.J. *Biochemistry of Sulfur*; Springer: Boston, MA, USA, 1986; ISBN 978-1-4757-9438-0. [\[Google Scholar\]](#)
3. Rendig, V.V.; Oputa, C.; McComb, E.A. Effects of sulfur deficiency on non-protein nitrogen, soluble sugars, and N/S ratios in young corn (*Zea mays* L.) plants. *Plant Soil* **1976**, *44*, 423–437. [\[Google Scholar\]](#) [\[CrossRef\]](#)
4. Reuveny, Z.; Dougall, D.K.; Trinity, P.M. Regulatory coupling of nitrate and sulfate assimilation pathways in cultured tobacco cells. *Proc. Natl. Acad. Sci. USA* **1980**, *77*, 6670–6672. [\[Google Scholar\]](#) [\[CrossRef\]](#) [\[PubMed\]](#)
5. Rennenberg, H. The fate of excess sulfur in higher plants. *Annu. Rev. Plant Physiol.* **1984**, *35*, 121–153. [\[Google Scholar\]](#) [\[CrossRef\]](#)
6. Andreae, M.O. Ocean-atmosphere interactions in the global biogeochemical sulfur cycle. *Mar. Chem.* **1990**, *30*, 1–29. [\[Google Scholar\]](#) [\[CrossRef\]](#)
7. Tolocka, M.P.; Turpin, B. Contribution of organosulfur compounds to organic aerosol mass. *Environ. Sci. Technol.* **2012**, *46*, 7978–7983. [\[Google Scholar\]](#) [\[CrossRef\]](#) [\[PubMed\]](#)
8. Mikkelsen, R.; Norton, R. Soil and fertilizer sulfur. *Better Crop.* **2013**, *97*, 7–9. [\[Google Scholar\]](#)
9. Chao, T.T.; Harward, M.E.; Fang, S.C. Cationic effects on sulfate adsorption by soils. *Soil Sci. Soc. Am. J.* **1963**, *27*, 35–38. [\[Google Scholar\]](#) [\[CrossRef\]](#)
10. Chao, T.T.; Harward, M.E.; Fang, S.C. Adsorption and desorption phenomena of sulfate ions in soils. *Soil Sci. Soc. Am. J.* **1962**, *26*, 234–237. [\[Google Scholar\]](#) [\[CrossRef\]](#)
11. Przygocka-Cyna, K.; Biber, M.; Przygocka-Cyna, K.; Grzebisz, W.; Pluta, M.; Grzebisz, W. Mineral density of onion bulbs as affected by fertilizers based on elemental sulfur. *J. Elem.* **2016**, *21*, 485–499. [\[Google Scholar\]](#) [\[CrossRef\]](#)
12. González-Morales, S.; Pérez-Labrada, F.; García-Enciso, E.L.; Leija-Martínez, P.; Medrano-Macías, J.; Dávila-Rangel, I.E.; Juárez-Maldonado, A.; Rivas-Martínez, E.N.; Benavides-Mendoza, A. Selenium and sulfur to produce *Allium* functional crops. *Molecules* **2017**, *22*, 558. [\[Google Scholar\]](#) [\[CrossRef\]](#)
13. Tea, I.; Genter, T.; Naulet, N.; Boyer, V.; Lummerzhim, M.; Kleiber, D. Effect of foliar sulfur and nitrogen fertilization on wheat storage protein composition and dough mixing properties. *Cereal Chem. J.* **2004**, *81*, 759–766. [\[Google Scholar\]](#) [\[CrossRef\]](#)
14. TSI Sulphur Fertilizer Types. Available online: <https://www.sulphurinstitute.org/fertilizer/sulphate.cfm> (accessed on 5 May 2019).
15. Johnson, D.W.; Cole, D.W. Anion mobility in soils: Relevance to nutrient transport from forest ecosystems. *Environ. Int.* **1980**, *3*, 79–90. [\[Google Scholar\]](#) [\[CrossRef\]](#)